# Report to the Mid-Atlantic Fishery Management Council: Fishery and Survey Data Updates Regarding the Northern Shortfin Squid (Illex illecebrosus) and Longfin Inshore Squid (Doryteuthis (Amerigo) pealeii) stocks through 2015 

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This report presents updates of the landings and survey relative abundance and biomass indices of the Northern shortfin squid (Illex illecebrosus) and longfin inshore squid (Doryteuthis (Amerigo) pealeii) stocks through 2015. This year, the data update report contains additional information requested by members of the Mid-Atlantic Fishery Management Council's (MAFMC) Scientific and Statistical Committee (SSC) to aid them in their task of providing an acceptable biological catch ( ABC ) recommendation for each of the two squid stocks. The additional information includes maps of the recent spatial distributions of the landings for both squid fisheries and stratified mean body weights from Northeast Fisheries Science Center (NEFSC) fall research bottom trawl surveys.

### 1.0 Illex illecebrosus

I. illecebrosus, a species with a lifespan of less than one year (Dawe and Beck 1997; Hendrickson 2004), constitutes a unit stock throughout its range in the Northwest Atlantic Ocean, from Southern Labrador to Florida (Dawe and Hendrickson, 1998). The northern stock component is located in Subareas 3 and 4 (Fig. 1.1) and is managed by the Northwest Atlantic Fisheries Organization (NAFO). The southern stock component is located in Subareas 5 and 6 between the waters of the Gulf of Maine and southern Florida and is managed by the MidAtlantic Fishery Management Council. Landings from the northern stock component have been very low (i.e., averaging < $7 \%$ of the total stock landings) since 1999 (Hendrickson and Showell 2013), and therefore, only the U. S. EEZ landings are presented herein.

### 1.1 Commercial Data

During 1963-1986, landings in Subareas 5 and 6 were predominately from international fleets and total landings averaged $11,027 \mathrm{mt}$ with a peak of $24,936 \mathrm{mt}$ in 1976 (Table 1.1, Fig. 1.2). Since 1987, landings of I. illecebrosus have been solely from a domestic smallmesh bottom trawl fishery which occurs primarily during June-October, when the species is available on the US continental shelf and upper slope. During 1987-2013, landings averaged $12,576 \mathrm{mt}$ and were characterized by two rise-and-fall periods. During 19872002, landings averaged $11,728 \mathrm{mt}$ and reached a peak of $23,568 \mathrm{mt}$ in 1998 when the
annual quota (TAC) of 19,000 mt was exceeded (Table 1.1, Fig. 1.1). During 2003-2013 landings averaged $13,810 \mathrm{mt}$ and reached a U.S fishery peak of $26,097 \mathrm{mt}$ in 2004 when the quota was exceeded again. In recent years, landings have been much lower; totaling $3,792 \mathrm{mt}$ in 2013 and $8,767 \mathrm{mt}$ in 2014. During 2015, landings declined further to 2,423 mt ; the lowest level since 2002, representing only $11 \%$ of the annual quota. During most years since 1996, a majority of the landings have been harvested by $6-15$ vessels (Arkhipkin et al. 2015), but only four vessels participated in the fishery intermittently during 2015. Low fishing effort was due to low prices of I. illecebrosus on the export market (G. Goodwin, pers. comm.). During 2013-2015, greater than $99 \%$ of the total stock landings were from the U.S. stock component.

Illex landings from the Dealer Weighout Database are considered the most accurate from 1996 onward because reporting of squid landings purchased by federally permitted dealers became mandatory in 1996. The submittal of Vessel Trip Reports (VTRs), which are the current source of fishing effort and location data, also became mandatory in 1996 for Illex moratorium permit holders.

The spatial distribution of I. illecebrosus landings by the directed fishery was mapped as the sum of the landings by ten-minute square (TNMS) for 2007-2010 and 2011-2014. The maps were prepared with data from trips with 1:1 matches between the Dealer Weighout Database and the VTR Database which were merged to create a combined database in accordance with the methods described in Wigley et al. (2008). The 2015 merged database had not been created as of the completion date of this report. The fishery was defined using the regulatory definition of a directed trip; trips with Illex landings > $4,536 \mathrm{~kg}$ ( $10,000 \mathrm{lbs}$ ). The data used to create the 2007-2010 and 2011-2014 maps represented $73 \%$ and $85 \%$, respectively, of the directed fishery landings. The distribution of landings for each of the two time blocks (Fig. 1.3) is typical for the fishery which is concentrated along the shelf edge in Statistical Areas 622, 626 and 632 (NEFSC 2006). The Illex fishery, which has no minimum codend mesh size requirement, is prohibited shoreward of 91 m during June through September to reduce the bycatch of D. pealeii. Based on the presence of TNMS with landings at depths < 91 m and $>400 \mathrm{~m}$, some misreporting of fishing location data is evident in both maps.

### 1.2 Survey Data

Indices of relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) were derived using data from the Northeast Fisheries Science Center (NEFSC) fall bottom trawl surveys conducted during 1967-2015 (Table 1.2, Fig. 1.4). Survey design and sampling protocols are described in Azarovitz (1981) and Politis et al. (2014). Catches from offshore strata 1-12 and 61-76 were used to compute the survey indices and the indices were standardized for gear and vessel changes that occurred during the time series (NEFSC 2006). The FSV H. B. Bigelow replaced the RV Albatross $I V$ as the primary survey vessel beginning in 2009. Indices for 2009 onward were converted to Albatross IV units using combined-season conversion factors, computed for all sizes combined (i.e., number per tow $=1.38$ and kg per tow $=1.41$ ), based on data from a 2008 vessel calibration study conducted with both vessels during the spring and
fall (Miller et al. 2010). Precision estimates (i.e., CVs computed as percentages) of the biomass and abundance indices account for the variance associated with these conversion factors (Table 1.2).

Relative abundance and biomass indices were highly variable during 1967-2015, as is typical for squid species given their sub-annual lifespans and the fact that spatial distribution patterns and recruitment are primarily determined by environmental factors (Boyle and Rodhouse 2005). Despite this variability, periods of high and low abundance are apparent. Periods of high abundance, during 1976-1981 and 1987-1990, were preceded by periods of low abundance during 1967-1974 and 1982-1986, respectively (Table 1.2, Fig. 1.4). During most years between 1991 and 2002, relative abundance fluctuated around the time series median of 8.0 squid per tow. Following a period of high variability, during 2003-2006, relative abundance declined to below the median in 2013 ( 4.7 squid per tow) then increased to 9.5 squid per tow in 2015 (Table 1.2, Fig. 1.4).

Stratified mean body weights of I. illecebrosus were computed as the annual stratified mean weight per tow divided by the stratified mean number per tow of squid caught during NEFSC fall research bottom trawl surveys. Mean body weight was highest during 1976-1981 (average $=295 \mathrm{~g}$ ) and were at or above the 1967-2014 median ( 123 g ) during 1967-1975, but have been much lower since 1982 (Hendrickson and Showell 2013). Since 2000, mean body weight has been consistently below the median and averaged 80 g during 2000-2014. In recent years, mean body weight declined from 86 g in 2011 to 55 g in 2015 (Fig. 1.5). Trends in squid mean body weight reflect the combined effects of growth, mortality, emigration and immigration. For I. illecebrosus, these factors are primarily influenced by large-scale oceanographic processes (Dawe et al. 2007). Of note is the fact that during the years when mean body weight was well below the median, annual landings during at least two of these years, 1998 and 2004, were the highest on record for the U. S. fishery (Fig. 1.5). I. illecebrosus is an oceanic squid species and a portion of the stock resides outside the range of NEFSC surveys; in deeper waters beyond 366 m , areas south of Cape Hatteras, and north of the Gulf of Maine in Subareas 3 and 4 (Hendrickson and Showell 2013; Arkhipkin et al. 2015). In addition, the survey bottom trawl gear may not sample all sizes of this semi-pelagic species efficiently (Hendrickson 2004). Therefore, NEFSC survey indices may represent a measure of the on-shelf availability of I. illecebrosus rather than a measure of relative abundance or biomass (NEFSC 2006).

### 1.3 References

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Table 1.1. USA EEZ (NAFO Subareas 5+6) landings (mt) of Illex illecebrosus, by fleet, during 1963-2015 and Total Allowable Catch (TAC, mt) during 1974-2015. The 2015 landings are preliminary and do not include all of the state data.

| Year | USA | International | Total | TAC ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1963 | 810 |  | 810 |  |
| 1964 | 358 | 2 | 360 |  |
| 1965 | 444 | 78 | 522 |  |
| 1966 | 452 | 118 | 570 |  |
| 1967 | 707 | 288 | 995 |  |
| 1968 | 678 | 2,593 | 3,271 |  |
| 1969 | 562 | 975 | 1,537 |  |
| 1970 | 408 | 2,418 | 2,826 |  |
| 1971 | 455 | 6,159 | 6,614 |  |
| 1972 | 472 | 17,169 | 17,641 |  |
| 1973 | 530 | 18,625 | 19,155 |  |
| 1974 | 148 | 20,480 | 20,628 | 71,000 |
| 1975 | 107 | 17,819 | 17,926 | 71,000 |
| 1976 | 229 | 24,707 | 24,936 | 30,000 |
| 1977 | 1,024 | 23,771 | 24,795 | 35,000 |
| 1978 | 385 | 17,207 | 17,592 | 30,000 |
| 1979 | 1,493 | 15,748 | 17,241 | 30,000 |
| 1980 | 299 | 17,529 | 17,828 | 30,000 |
| 1981 | 615 | 14,956 | 15,571 | 30,000 |
| 1982 | 5,871 | 12,762 | 18,633 | 30,000 |
| 1983 | 9,775 | 1,809 | 11,584 | 30,000 |
| 1984 | 9,343 | 576 | 9,919 | 30,000 |
| 1985 | 5,033 | 1,082 | 6,115 | 30,000 |
| 1986 | 6,493 | 977 | 7,470 | 30,000 |
| 1987 | 10,102 | 0 | 10,102 | 30,000 |
| 1988 | 1,958 | 0 | 1,958 | 30,000 |
| 1989 | 6,801 | 0 | 6,801 | 30,000 |
| 1990 | 11,670 | 0 | 11,670 | 30,000 |
| 1991 | 11,908 | 0 | 11,908 | 30,000 |
| 1992 | 17,827 | 0 | 17,827 | 30,000 |
| 1993 | 18,012 | 0 | 18,012 | 30,000 |
| 1994 | 18,350 | 0 | 18,350 | 30,000 |
| 1995 | 13,976 | 0 | 13,976 | 30,000 |
| 1996 | 16,969 | 0 | 16,969 | 21,000 |
| 1997 | 13,356 | 0 | 13,356 | 19,000 |
| 1998 | 23,568 | 0 | 23,568 | 19,000 |
| 1999 | 7,388 | 0 | 7,388 | 19,000 |
| 2000 | 9,011 | 0 | 9,011 | 24,000 |
| 2001 | 4,009 | 0 | 4,009 | 24,000 |
| 2002 | 2,750 | 0 | 2,750 | 24,000 |
| 2003 | 6,391 | 0 | 6,391 | 24,000 |

Table 1.1. (cont.)

| Year | USA | International | Total | TAC $^{\mathbf{1}}$ |
| :---: | ---: | ---: | ---: | ---: |
| 2004 | 26,097 | 0 | 26,097 | 24,000 |
| 2005 | 12,011 | 0 | 12,011 | 24,000 |
| 2006 | 13,944 | 0 | 13,944 | 24,000 |
| 2007 | 9,022 | 0 | 9,022 | 24,000 |
| 2008 | 15,900 | 0 | 15,900 | 24,000 |
| 2009 | 18,418 | 0 | 18,418 | 24,000 |
| 2010 | 15,825 | 0 | 15,825 | 24,000 |
| 2011 | 18,797 | 0 | 18,797 | 23,328 |
| 2012 | 11,709 | 0 | 11,709 | 22,915 |
| 2013 | 3,792 | 0 | 3,792 | 22,915 |
| 2014 | 8,767 | 0 | 8,767 | 22,915 |
| 2015 | 2,423 | 0 | 2,423 | 22,915 |
| Averages |  |  |  |  |
| $1963-1986$ | 1,950 | 9,472 | 11,027 |  |
| $1987-2014$ | 12,440 | 0 | 12,440 |  |
| $1963-2014$ | 7,598 | 4,272 | 11,788 |  |

${ }^{1}$ TACs during 1974 and 1975 are for Illex illecebrosus and Doryteuthis pealeii combined.

Table 1.2. Illex illecebrosus relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) indices, and CVs (\%), derived from NEFSC fall bottom trawl surveys (offshore strata 1-12 and 61-76) conducted during 1967-2015. FSV H. B. Bigelow indices for 2009 onward were converted to RV Albatross IV units using combined-season conversion factors computed for all sizes combined based on Miller et al. (2010). CVs from 2009 onward account for the variance associated with the FSV Bigelow conversion factors.

| Year | Number per tow | CV (\%) | Kg per tow | CV (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 1967 | 1.6 | 17 | 0.24 | 17 |
| 1968 | 1.6 | 21 | 0.31 | 17 |
| 1969 | 0.6 | 23 | 0.07 | 26 |
| 1970 | 2.3 | 21 | 0.27 | 15 |
| 1971 | 1.7 | 12 | 0.34 | 14 |
| 1972 | 2.2 | 25 | 0.29 | 15 |
| 1973 | 1.5 | 24 | 0.35 | 25 |
| 1974 | 2.8 | 40 | 0.39 | 30 |
| 1975 | 8.7 | 36 | 1.42 | 18 |
| 1976 | 20.6 | 16 | 7.02 | 19 |
| 1977 | 12.6 | 18 | 3.74 | 18 |
| 1978 | 19.3 | 21 | 4.53 | 26 |
| 1979 | 19.4 | 11 | 6.05 | 11 |
| 1980 | 13.8 | 15 | 3.29 | 18 |
| 1981 | 27.1 | 32 | 9.34 | 40 |
| 1982 | 3.9 | 15 | 0.60 | 13 |
| 1983 | 1.7 | 14 | 0.23 | 13 |
| 1984 | 4.5 | 17 | 0.52 | 19 |
| 1985 | 2.4 | 17 | 0.36 | 18 |
| 1986 | 2.1 | 15 | 0.26 | 17 |
| 1987 | 15.8 | 31 | 1.53 | 29 |
| 1988 | 23.2 | 25 | 3.00 | 24 |
| 1989 | 22.4 | 45 | 3.31 | 57 |
| 1990 | 16.6 | 12 | 2.40 | 13 |
| 1991 | 5.2 | 17 | 0.69 | 18 |
| 1992 | 8.2 | 15 | 0.80 | 16 |
| 1993 | 10.4 | 19 | 1.60 | 20 |
| 1994 | 6.8 | 24 | 0.86 | 25 |
| 1995 | 8.0 | 30 | 0.70 | 39 |
| 1996 | 10.8 | 22 | 0.93 | 19 |
| 1997 | 5.8 | 25 | 0.52 | 17 |
| 1998 | 14.6 | 29 | 1.40 | 50 |
| 1999 | 1.4 | 16 | 0.19 | 17 |
| 2000 | 4.5 | 28 | 0.71 | 22 |
| 2001 | 6.4 | 27 | 0.32 | 23 |
| 2002 | 28.5 | 61 | 0.44 | 19 |
| 2003 |  | 1.95 | 67 |  |
|  |  |  |  |  |


| Table 1.2 (cont.) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Number per tow | CV (\%) | Kg per tow | CV (\%) |
| 2004 | 5.1 | 24 | 0.41 | 22 |
| 2005 | 11.0 | 35 | 0.74 | 41 |
| 2006 | 29.5 | 43 | 2.85 | 31 |
| 2007 | 15.7 | 33 | 1.31 | 33 |
| 2008 | 10.4 | 22 | 0.98 | 20 |
| 2009 | 8.7 | 18 | 0.93 | 21 |
| 2010 | 10.0 | 23 | 0.53 | 23 |
| 2011 | 6.3 | 20 | 0.54 | 20 |
| 2012 | 8.0 | 17 | 0.54 | 15 |
| 2013 | 4.7 | 17 | 0.36 | 16 |
| 2014 | 8.3 | 14 | 0.64 | 14 |
| 2015 | 9.5 | 36 | 0.52 | 16 |
| Median |  |  |  |  |
| $1967-2014$ | 8.0 |  | 0.70 |  |



Figure 1.1. Northwest Atlantic Fisheries Organization (NAFO) nominal catch reporting areas, Subareas 3-6 and associated Divisions, for fisheries occurring in the Northwest Atlantic Ocean.


Figure 1.2 Landings ( 000 's mt) of Illex illecebrosus from the USA EEZ (NAFO Subareas 5+6), by fleet, and TACs ( 000 's mt ) during 1963-2015. The 2015 landings are preliminary.


Figure 1.3. Distribution of landings ( mt ) from bottom trawl trips with Illex illecebrosus landings > 4.536 mt ( $10,000 \mathrm{lbs}$ ), by ten-minute square, during 2007-2010 and 2011-2014. The Southern Gear Restricted Area (GRA) is in effect from January 1 to March 15 and the Northern GRA is in effect from November 1 to December 31. Squid fishing does not occur in the GRAs during these time periods because bottom trawls with codend mesh sizes $<127 \mathrm{~mm}$ diamond mesh ( 5.0 in ., inside stretched mesh) are prohibited. East of $72^{\circ} 30^{\prime} \mathrm{N}$, squid fishing is only permitted in small-mesh exemption areas which are not shown in the figure.


Figure 1.4. Illex illecebrosus indices of relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) derived from NEFSC fall bottom trawl survey data, 1967-2015.


Figure 1.5. Stratified mean body weights of Illex illecebrosus (stratified mean kg per tow / stratified mean number per tow) derived from NEFSC fall bottom trawl survey data, 1967-2015. The dashed line represents the 1967-2014 median body weight.

### 2.0 Doryteuthis (Amerigo) pealeii

Longfin inshore squid (Doryteuthis (Amerigo) pealeii), hereafter referred to as $D$. pealeii, also has a lifespan of less than one year (Macy and Brodziak 2001). D. pealeii inhabits continental shelf and upper slope waters between southern Newfoundland and the Gulf of Venezuela, including the Gulf of Mexico and the Caribbean Sea (Jereb et al. 2010). The species is most abundant between Georges Bank and Cape Hatteras, North Carolina where a small-mesh bottom trawl fishery occurs throughout the year. The U.S. East coast longfin squid population is managed as a single stock based on evidence from genetics studies (Arkhipkin et al. 2015).

### 2.1 Commercial Data

During 1967-1984, landings of D. pealeii in Subareas 5 and 6 were predominately from international fleets. During 1963-1986, total landings averaged 16,489 mt with a peak of 37,613 mt in 1973 (Table 2.1, Fig. 2.1). Since 1987, landings of D. pealeii have been solely from a domestic, small-mesh bottom trawl fishery which occurs throughout the year. During 1987-2014, landings averaged $15,484 \mathrm{mt}$ with a peak of $23,738 \mathrm{mt}$ in 1989 (Table 2.1, Fig. 2.1). During 1999-2010, landings gradually declined from $19,173 \mathrm{mt}$ to $6,913 \mathrm{mt}$, respectively, then increased again to $12,820 \mathrm{mt}$ in 2012. During 2013-2015, landings were slightly above the 2007-2014 average ( $10,689 \mathrm{mt}$ ). Preliminary landings for 2015, which do not include all state landings, totaled $11,928 \mathrm{mt}$. Annual quotas (TACs) for the U.S. fishery were exceeded only once, during 2000, but landings totaled $94 \%-99 \%$ of the annual quotas during 2002, 2005 and 2006 when the quota was only $17,000 \mathrm{mt}$ (Table 2.1). During 20092015, the quota was increased and ranged between $18,667 \mathrm{mt}$ in 2010 and 22,445 mt in 2015 and about $50 \%$ of the quota was harvested on average.

In addition to stock abundance, availability, and ex-vessel price of $D$. pealeii, monthly landings trends have been affected by in-season longfin squid quotas since 2000. Landings trends have also been affected by trimester-based Atlantic butterfish (Peprilus triacanthus) catch quotas since 2011 and discard quotas since March 5, 2013. One or more longfin squid fishery closures have occurred per year, with the exception of 2010, 2013, and 2015, due to attainment of in-season quota buffers for $D$. pealeii (Table 2.2). D. pealeii landings were more evenly distributed across months during 1987-1995 than thereafter (Fig. 2.2). Monthly landings trends were similar during 1996-1999 (no in-season quotas) and 2001-2006 (quarterly quotas) and generally declined from a peak in February (15-16\% of annual landings) to a minimum of $2 \%$ in June then increased to $13 \%$ during October (Arkhipkin et al. 2015). During the current management regime of trimester-based quotas (2000 and 20072014), the landings peak shifted to October (13\%) with a secondary peak of $12 \%$ harvested during July. This shift in the landings peakis in part attributable to a consistent decline between 1998 and 2012 in the number of large vessels involved in the offshore fishery during November-April (Arkhipkin et al. 2015). Landings during Trimesters 1, 2 and 3 totaled 32\%, $37 \%$, and $32 \%$, respectively, of the combined landings for the period including 2000 and 2007-2014, but the interannual variability of landings by trimester was high (Fig. 2.3). The current quota allocations for Trimesters $1-3$ are $43 \%, 17 \%$, and $40 \%$, respectively. As of 2010, a Trimester 2 quota increase of up to $150 \%$ is allowable when a comparable underage of the Trimester 1 quota occurs.
D. pealeii landings from 1996 onward are considered the most accurate because reporting of squid landings purchased by federally permitted dealers became mandatory in 1996. The submittal of Vessel Trip Reports (VTRs), which are the current source of fishing effort and location data, also became mandatory in 1996 for longfin squid/butterfish moratorium permit holders.

The spatial distribution of $D$. peleaii landings by the directed fishery was mapped using the same methodology as described above in Section 1.1 but separate maps were prepared for each trimester. The fishery was defined using the regulatory definition of a directed trip; trips with Illex landings > 1,134 kg ( $2,500 \mathrm{lbs}$ ). The mapped landings data for each trimester represented $69 \%, 72 \%$ and $74 \%$, respectively, of the 2007-2010 directed fishery landings and $88 \%, 90 \%$ and $87 \%$, respectively, and of the 2011-2014 directed fishery landings. The landings distribution patterns indicated in Figs. 2.4 and 2.5 are typical for the fishery, which generally occurs offshore during October-March and inshore during April-September (NEFSC 2011). The Trimester 2 map for 2011-2014 (Fig. 2.4) reflects the unusually high landings that occurred inshore during 2012 (totaling $62 \%$ of the annual landings) and which resulted in increased incidental catches of longfin squid in the offshore Illex fishery. Based on the presence of TNMS with landings at depths $>400 \mathrm{~m}$, some misreporting of fishing location data is evident in all six maps.

### 2.2 Survey Data

The efficiency of the NEFSC survey bottom trawl gear for $D$. pealeii, a diel vertical migrator, is highest during the daytime (Jacobson et al. 2015). Indices of relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) were derived using daytime tows (tows with solar zenith angles of $43^{\circ}-80^{\circ}$ ) from NEFSC fall bottom trawl surveys conducted during 1975-2015 based on the methods used in the most recent stock assessment (NEFSC 2011). Indices include catches from inshore strata 5, 8, 11, 14, 17, 20, 23, 26, 29, 32, 35, 38, 41, 44-46, 59-61 and 65-66 plus offshore strata 1-23, 25-26, and 61-76. FSV H. B. Bigelow indices for 2009 onward were converted to RV Albatross IV units using daytime conversion factors, computed for all sizes combined, from a 2008 vessel calibration study conducted with both vessels during the fall (NEFSC 2011). Coefficients of variation of the survey indices account for the variance associated with the Bigelow conversion factors (Table 2.3).

Since 2009, NEFSC survey strata ranging in depth from 6 m to 18 m can no longer be sampled because they are too shallow for operation of the FSV H. B. Bigelow. These strata constitute habitat for longfin squid, primarily squid $\leq 10 \mathrm{~cm}$ dorsal mantle length, during the fall (Brodziak and Hendrickson 1999). Therefore, longfin squid relative abundance and biomass indices are also provided for the 2007-2015 fall Northeast Area Monitoring and Assessment Program (NEAMAP) research bottom trawl surveys. The NEAMAP surveys, which are funded by NOAA Fisheries and conducted by staffs from the Virginia Institute of Marine Science, were designed to sample the nearshore strata formerly sampled during NEFSC surveys. The stratified random survey is conducted is conducted by a fishing vessel (FV Darana R) with a smaller version of the trawl gear used by the FSV Bigelow. The
difference between the NEAMAP and NEFSC survey gears is a 3-in. cookie sweep versus a 16 -in. diameter rockhopper sweep (14-in. diameter on the wings, Politis et al. 2014), respectively. Both surveys use the same towing protocols ( 20 min . at 3.0 kts ). During the fall NEAMAP surveys, most of the tows were conducted during October. The NEAMAP surveys occur between Rhode Island Sound and Cape Hatteras, North Carolina and all stations are sampled between sunrise and sunset (approximately 0715-1715). Survey sampling methods and protocols and the methods used to derive the longfin squid relative abundance and biomass indices are provided in Bonzek et al. (2015). However, the longfin squid survey indices included in Bonzek et al. (2015) contained some errors. Therefore, the corrected fall indices are provided herein. The indices were derived by NEAMAP staff as the backtransformed, stratified geometric mean catch per area-swept (in numbers and kg ) and include longfin squid catches from all survey strata (Jim Gartland, pers. comm.).

Relative abundance and biomass indices of D. pealeii (Fig. 2.6) were much more variable than the indices for I. illecebrosus (Figure 1.3), making any underlying trends difficult to discern. In recent years, relative abundance declined from the third highest point in the time series during 2006 ( 1,778 squid per tow) to 416 squid per tow in 2009 , a level below the time series median of 651 squid per tow (Table 2.3, Fig. 2.6). During 2012-2014 relative abundance was above the median, but declined from 1,371 squid per tow to 744 squid per tow, respectively, then dropped below the median to 596 squid per tow in 2015.

NEAMAP relative abundance indices increased from below the 2007-2014 median (79 squid per tow) in 2010 ( 30 squid per tow) to a peak of 208 squid per tow in 2014 (Table 2.4, Fig. 2.7). During 2015, relative abundance declined but remained above the median and was slightly greater than the 2013 index. Trends in the relative biomass indices were similar to trends in relative abundance. The NEAMAP fall survey indices were provided as backtransformed geometric means, and as such, cannot be compared directly with the NEFSC fall survey indices which were computed as arithmetic means. In addition, there are no gear/vessel conversion coefficients available with which to standardize longfin squid catches between the two surveys and longfin squid migration rates between the two survey areas during the fall are unknown.

Stratified mean body weights of $D$. pealeii were computed, as described above in Section 1.2, using data from NEFSC fall research bottom trawl surveys. During most years between 1976 and 1994, mean body weights were above the 1975-2014 median of 20 g and were below the median during most years thereafter (Fig. 2.8). Trends in squid mean body weight reflect the combined effects of growth, mortality, emigration and immigration. For D. pealeii, these factors are primarily influenced by water temperatures during the early life history stages (Dawe et al. 2007).

### 2.3 References

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Table 2.1. Landings (mt) of Doryteuthis pealeii, by fleet, during 1963-2015. The 2015 landings are preliminary and do not include all of the state data.

| Year | USA | International | Total | TAC ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1963 | 1,294 | 0 | 1,294 |  |
| 1964 | 576 | 2 | 578 |  |
| 1965 | 709 | 99 | 808 |  |
| 1966 | 722 | 226 | 948 |  |
| 1967 | 547 | 1,130 | 1,677 |  |
| 1968 | 1,084 | 2,327 | 3,411 |  |
| 1969 | 899 | 8,643 | 9,542 |  |
| 1970 | 653 | 16,732 | 17,385 |  |
| 1971 | 727 | 17,442 | 18,169 |  |
| 1972 | 725 | 29,009 | 29,734 |  |
| 1973 | 1,105 | 36,508 | 37,613 |  |
| 1974 | 2,274 | 32,576 | 34,850 | 70,000 |
| 1975 | 1,621 | 32,180 | 33,801 | 70,000 |
| 1976 | 3,602 | 21,682 | 25,284 | 44,000 |
| 1977 | 1,088 | 15,586 | 16,674 | 44,000 |
| 1978 | 1,476 | 9,355 | 10,831 | 44,000 |
| 1979 | 4,252 | 13,068 | 17,320 | 44,000 |
| 1980 | 3,996 | 19,750 | 23,746 | 44,000 |
| 1981 | 2,316 | 20,212 | 22,528 | 44,000 |
| 1982 | 2,848 | 15,805 | 18,653 | 44,000 |
| 1983 | 10,867 | 11,720 | 22,587 | 44,000 |
| 1984 | 7,689 | 11,031 | 18,720 | 44,000 |
| 1985 | 6,899 | 6,549 | 13,448 | 44,000 |
| 1986 | 11,525 | 4,598 | 16,123 | 44,000 |
| 1987 | 10,367 | 2 | 10,369 | 44,000 |
| 1988 | 18,593 | 3 | 18,596 | 44,000 |
| 1989 | 23,733 | 5 | 23,738 | 44,000 |
| 1990 | 15,399 | 0 | 15,399 | 44,000 |
| 1991 | 20,299 | 0 | 20,299 | 44,000 |
| 1992 | 19,018 | 0 | 19,018 | 44,000 |
| 1993 | 23,020 | 0 | 23,020 | 44,000 |
| 1994 | 23,480 | 0 | 23,480 | 44,000 |
| 1995 | 18,880 | 0 | 18,880 | 36,000 |
| 1996 | 12,503 | 0 | 12,503 | 25,000 |
| 1997 | 16,270 | 0 | 16,270 | 21,000 |
| 1998 | 19,145 | 0 | 19,145 | 21,000 |
| 1999 | 19,173 | 0 | 19,173 | 21,000 |
| 2000 | 17,540 | 0 | 17,540 | 15,000 |
| 2001 | 14,345 | 0 | 14,345 | 17,000 |
| 2002 | 16,868 | 0 | 16,868 | 17,000 |
| 2003 | 11,941 | 0 | 11,941 | 17,000 |


| Table 2.1 (cont.) |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: |
| Year | USA | International | Total | TAC |
| 2004 | 14,800 |  |  | 17,000 |
| 2005 | 16,724 | 0 | 14,800 | 17,000 |
| 2006 | 15,928 | 0 | 16,724 | 17,000 |
| 2007 | 12,354 | 0 | 15,928 | 12,354 |
| 2008 | 11,406 | 0 | 11,406 | 17,000 |
| 2009 | 9,307 | 0 | 9,307 | 19,000 |
| 2010 | 6,913 | 0 | 6,913 | 18,667 |
| 2011 | 9,556 | 0 | 9,556 | 19,906 |
| 2012 | 12,820 | 0 | 12,820 | 22,220 |
| 2013 | 11,183 | 0 | 11,183 | 22,049 |
| 2014 | 12,063 | 0 | 12,063 | 22,049 |
| 2015 | 11,928 | 0 | 11,928 | 22,445 |
| Averages |  |  |  |  |
| $1963-1986$ | 2,896 | 14,184 | 16,489 |  |
| $1987-2014$ | 15,484 | 0 | 15,484 |  |
| $2007-2014$ | 10,689 | 0 | 10,689 |  |
| $1963-2014$ | 9,674 | 6,397 | 15,948 |  |

TACs during 1974 and 1975 are for Illex illecebrosus and Doryteuthis pealeii combined.

Table 2.2. Doryteuthis pealeii fishery closures which occurred during 2000-2015 when in-season quotas were in effect. Quotas were trimester-based during 2000 and 2007-2015 and quarterly during 2001-2006. Closures trigger a regulatory trip limit of $1,134 \mathrm{~kg}$ of $D$. pealeii.

| Year | Quota period I | Quota period II | Quota period III | Quota period IV |
| :--- | ---: | ---: | ---: | ---: |
| 2000 | Mar 25 - Apr 30 | Jul 1- Aug 31 | Sep 7 - Oct 6, <br> Oct 26 - Dec 31 |  |
| 2001 |  | May 29 - Jun 30 <br> May 28 - Jun 30 | Aug 16 - Sep 30 | Nov 2 - Dec 11, <br> Dec 24 - Dec 31 |
| 2002 |  |  |  |  |
| 2003 | Mar 25 - Mar 31 |  |  | Dec 18 - Dec 31 |
| 2004 | Mar 5 - Mar 31 | Apr 25 - Jun 30 |  |  |
| 2005 | Feb 20 - Mar 31 | Apr 21 - Apr 27, | Sep 2 - Sep 30 |  |
| 2006 | Feb 13 - Mar 31 | May 23 - Jun 30 |  |  |
| 2007 | Apr 13 - Apr 30 |  |  |  |
| 2008 |  | July 17 - Aug 31 |  |  |
| 2009 |  | Aug 6 - Aug 31 |  |  |
| 2010 |  | Aug 23-Aug 31 |  |  |
| 2011 |  | July 10 - Aug 31 |  |  |
| 2012 | *Apr 17-Apr 30 | Aug 11-Aug 31 |  |  |
| 2013 |  |  |  |  |
| 2014 |  |  |  |  |
| 2015 |  |  |  |  |

* Closure due to reaching incidental catch cap for Atlantic butterfish (Peprilus triacanthus).

Table 2.3. Relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) indices for Doryteuthis pealeii, and CVs (\%), derived using daytime tows (tows conducted when solar zenith angles were between $43^{\circ}$ and $80^{\circ}$ ) from NEFSC fall bottom trawl surveys conducted during 1975-2015. Indices include catches from inshore strata 5, 8, 11, 14, 17, 20, 23, 26, $29,32,35,38,41,44-46,59-61$ and 65-66 plus offshore strata 1-23, 25-26, and 61-76. FSV H. B. Bigelow indices for 2009 onward were converted to RV Albatross IV units using "daytime", fall survey conversion factors computed for all sizes combined based on NEFSC (2011). CVs from 2009 onward account for the variance associated with the FSV Bigelow conversion factors.

| Year | Number per tow | CV (\%) | Kg per tow | CV (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 1975 | 1,038 | 14 | 14.4 | 11 |
| 1976 | 730 | 12 | 18.8 | 15 |
| 1977 | 513 | 14 | 11.5 | 18 |
| 1978 | 270 | 16 | 7.6 | 11 |
| 1979 | 376 | 12 | 8.2 | 12 |
| 1980 | 562 | 13 | 14.2 | 8 |
| 1981 | 402 | 10 | 12.5 | 6 |
| 1982 | 529 | 13 | 12.4 | 15 |
| 1983 | 814 | 14 | 23.7 | 20 |
| 1984 | 625 | 10 | 20.8 | 17 |
| 1985 | 709 | 15 | 19.6 | 11 |
| 1986 | 720 | 13 | 14.8 | 4 |
| 1987 | 101 | 9 | 2.8 | 9 |
| 1988 | 651 | 14 | 9.3 | 13 |
| 1989 | 830 | 25 | 21.5 | 34 |
| 1990 | 480 | 12 | 10.4 | 14 |
| 1991 | 375 | 12 | 11.5 | 10 |
| 1992 | 2,029 | 27 | 10.4 | 20 |
| 1993 | 185 | 26 | 4.9 | 10 |
| 1994 | 905 | 11 | 27.5 | 15 |
| 1995 | 340 | 15 | 5.8 | 8 |
| 1996 | 306 | 18 | 3.8 | 20 |
| 1997 | 548 | 21 | 10.3 | 22 |
| 1998 | 381 | 14 | 5.3 | 14 |
| 1999 | 1,341 | 10 | 15.4 | 10 |
| 2000 | 1,035 | 12 | 30.4 | 7 |
| 2001 | 431 | 11 | 8.5 | 8 |
| 2002 | 1,960 | 4 | 23.4 | 5 |
| 2003 | 951 | 8 | 14.0 | 11 |
| 2004 | 1,055 | 14 | 8.6 | 10 |
| 2005 | 530 | 14 | 9.9 | 20 |
| 2006 | 1,778 | 10 | 22.9 | 6 |
| 2007 | 1,111 | 17 | 10.1 | 18 |
| 2008 | 667 | 18 | 11.3 | 25 |
| 2009 | 416 | 9 | 6.3 | 13 |
| 2010 | 708 | 21 | 15.0 | 12 |


| Table 2.3 (cont.) |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Year | Number per tow | CV (\%) | Kg per tow | CV (\%) |
| 2011 | 339 | 16 | 6.6 | 12 |
| 2012 | 1,371 | 11 | 21.0 | 9 |
| 2013 | 1,012 | 36 | 13.3 | 16 |
| 2014 | 744 | 13 | 12.8 | 11 |
| 2015 | 596 | 8 | 9.7 | 8 |
| Median |  |  |  |  |
| $1975-2014$ | 659 |  | 12.0 |  |

Table 2.4 Relative abundance and biomass indices (back-transformed, stratified geometric mean number and kg, respectively, per area-swept) for Doryteuthis pealeii, and $95 \%$ confidence limits, derived using data from NEAMAP fall bottom trawl surveys conducted during 2007-2015. All NEAMAP survey tows are conducted during the daytime, between sunrise and sunset.

| Year | LCI | Number per tow | UCI | LCI | Kg per tow | UCI |
| ---: | ---: | :---: | ---: | :---: | :---: | :---: |
| 2007 | 116.5 | 143 | 174.4 | 4.16 | 4.9 | 5.84 |
| 2008 | 38.1 | 48 | 60.9 | 2.40 | 2.8 | 3.33 |
| 2009 | 91.6 | 118 | 152.0 | 4.95 | 5.8 | 6.85 |
| 2010 | 29.3 | 38 | 48.8 | 2.88 | 3.4 | 4.08 |
| 2011 | 38.1 | 46 | 56.0 | 2.67 | 3.1 | 3.47 |
| 2012 | 49.1 | 60 | 74.0 | 2.90 | 3.4 | 3.92 |
| 2013 | 76.5 | 97 | 123.6 | 4.76 | 5.7 | 6.75 |
| 2014 | 208.4 | 260 | 324.0 | 9.44 | 11.0 | 12.79 |
| 2015 | 87.9 | 105 | 126.3 | 6.48 | 7.5 | 8.69 |



Figure 2.1. Landings ( 000 's mt) of Doryteuthis pealeii, by fleet, and TACs ( 000 's mt) during 1963-2015. The 2015 landings are preliminary and do not include all state data.


Figure 2.2. Landings (\%) of Doryteuthis pealeii, by month, during four fishery management periods: annual quotas without (1987-1995) and with (1996-1999) mandatory landings reporting; quarterly quotas (2001-2006); and trimester quotas (2000 and 2007-2015) in comparison to preliminary landings for 2015.


Figure 2.3. Landings (\%) of Doryteuthis pealeii, by trimester (T1, T2 and T3), during the current management regime of trimester quotas (2007-2015). Landings during 2015 are preliminary. Asterisks indicate $D$. pealeii fishery closures during a portion of the trimester indicated due to attaining the buffer percentage of the specific trimester quota for $D$. pealeii. The fishery closure that occurred during T1 was attributable to attaining the incidental catch cap for Atlantic butterfish (Peprilus triancanthus).


Figure 2.4. Distribution of landings (mt) from bottom trawl trips with Doryteuthis pealeii landings > $1.134 \mathrm{mt}(2,500 \mathrm{lbs})$, by trimester and ten-minute square, during 2007-2010. The Southern Gear Restricted Area (GRA) is in effect from January 1 to March 15 (Trimester 1) and the Northern GRA is in effect from November 1 to December 31. Squid fishing does not occur in these GRAs when they are in effect because bottom trawls with a codend mesh size < 127 mm diamond mesh ( 5.0 in ., inside stretched mesh) are prohibited. East of $72^{\circ} 30^{\circ} \mathrm{N}$, squid fishing is only permitted in small-mesh exemption areas.


Figure 2.5. Distribution of landings ( mt ) from bottom trawl trips with Doryteuthis pealeii landings $>1.134 \mathrm{mt}(2,500 \mathrm{lbs})$, by trimester and ten-minute square, during 2011-2014. The Southern Gear Restricted Area (GRA) is in effect from January 1 to March 15 (Trimester 1) and the Northern GRA is in effect from November 1 to December 31. Squid fishing does not occur in these GRAs when they are in effect because bottom trawls with a codend mesh size < 127 mm diamond mesh (5.0 in., inside stretched mesh) are prohibited. East of $72^{\circ} 30^{\circ} \mathrm{N}$, squid fishing is only permitted in small-mesh exemption areas.


Figure 2.6. Doryteuthis pealeii indices of relative abundance (stratified mean number per tow) and biomass (stratified mean kg per tow) derived using daytime tows (solar zenith angles of $43^{\circ}$ $80^{\circ}$ ) from NEFSC fall bottom trawl surveys conducted during 1975-2015.


Figure 2.7. Doryteuthis pealeii indices of relative abundance (stratified geometric mean number per tow) and biomass (stratified geometric mean kg per tow) derived using catch data from NEAMAP fall bottom trawl surveys conducted during 2007-2015.


Figure 2.8. Stratified mean body weights of Doryteuthis pealeii (stratified mean kg per tow / stratified mean number per tow) derived from NEFSC fall bottom trawl survey data, 1967-2015. The dashed line represents the 1967-2014 median body weight.

